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**Effects of bison on willow and cottonwood
in northern Yellowstone National Park**

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13 ABSTRACT

14 On the northern ungulate winter range of Yellowstone Park, willow (*Salix* spp.) and
15 cottonwood (*Populus angustifolia* and *P. balsamifera*) have increased in height and cover in
16 some places since the reintroduction of wolves (*Canis lupus*) and the subsequent changes in elk
17 (*Cervus elaphus*) behavior and population densities. However, in the Lamar Valley, an important
18 part of this winter range, many plants are still intensively browsed and recruitment has been
19 limited. As elk numbers have declined and their distribution has changed in recent years, bison
20 (*Bison bison*) have increased on the northern range. To distinguish bison effects from those of
21 elk, we measured browsing that occurred in summer. We found average summer browse rates of
22 84% for willow and 54% for cottonwood seedlings in the summer of 2010, demonstrating that
23 bison have become significant browsers in the Lamar Valley. Plants were increasing in size
24 except where intensively browsed by bison, suggesting that a release from elk browsing has
25 occurred, and that a trophic cascade is occurring from wolves to plants, mediated by both elk and
26 bison. Release of bison from competition with elk, low levels of predation on bison, and lack of
27 opportunity for migration and range expansion may be factors contributing to a high
28 concentration of bison, with resulting effects on plant communities and biodiversity.

29

30 **Key words:** bison, browsing, trophic cascade, wolves, elk, Lamar Valley

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31

32 1. INTRODUCTION

33 Some ecosystems of western North America were shaped in the past by bison (*Bison*
34 *bison*) and the ecological effects of these iconic animals may again be a factor, with recent
35 efforts to restore them to portions of their former range (Sanderson et al., 2008; Gates et al.,
36 2010). In Yellowstone National Park, elk (*Cervus elaphus*) numbers have decreased following
37 wolf (*Canis lupus*) reintroductions in 1995 and 1996, but the bison population has continued to
38 grow, maintained below a peak of about 5000 by large-scale culling when bison leave the park in
39 winter (White and Garrott, 2005; Plumb et al., 2009; White et al., 2010; 2011). White et al.
40 (2011) reported that current management practices, in which bison are kept close to park
41 boundaries in winter and hazed back into the park in early spring, are likely to lead to high
42 population densities and density-dependence among bison, possibly causing deterioration of
43 range resources and ecological processes.

44 Valleys in the northern part of Yellowstone National Park are used as winter range by
45 elk, bison, and other ungulates (Figure 1; Singer and Norland, 1994). In this area, called the
46 northern ungulate winter range, or “northern range,” willow (*Salix* spp.), cottonwood (*Populus*
47 *angustifolia* and *P. balsamifera*) and aspen (*Populus tremuloides*) declined in the 20th century,
48 primarily due to browsing by elk in winter (Kay, 1994; Chadde and Kay, 1996; Singer, 1996;
49 Keigley, 1997, 2000; Romme et al., 2001; National Research Council, 2002; Barmore, 2003;
50 Singer et al., 2003; Beyer, 2006; Wagner, 2006). Beaver (*Castor canadensis*), which depend on
51 these plants, also declined in number and range resulting in loss of wetlands and further decline
52 of willows (Wolf et al., 2007; Bilyeu et al., 2008; Smith and Tyers, 2008). Since the return of
53 wolves to the northern range, elk population size, spatial distribution, and foraging behavior have
54 changed (Laundre et al., 2001; Hernandez and Laundre, 2005; White et al., 2010; White et al., in
55 press). Probably as a result of these changes, woody browse plants have increased in height and
56 cover in some places (Ripple and Beschta, 2006; Beschta and Ripple, 2007; Beyer et al., 2007;
57 Ripple and Beschta, in press), and beaver have increased in number and range (Smith et al.,
58 2003; Smith and Tyers, 2008). For example, few cottonwood trees grew to maturity on the
59 northern range after the early 20th century (Beschta, 2005), and cottonwood saplings were kept
60 short (<1 m) by browsing (Keigley, 1997; Beschta, 2003). Between 2001 and 2006 cottonwoods
61 again began to grow tall enough to begin to escape elk browsing (>2 m) in places along the east

62 edge of the Lamar Valley, on an island in the Lamar River, and on Soda Butte Creek (Ripple and
63 Beschta, 2003; Beschta and Ripple, 2010). Willows also increased in height in some places
64 (Ripple and Beschta, 2006; Beschta and Ripple, 2007). However, in most of the Lamar Valley,
65 west of the Soda Butte Creek confluence (Figure 1), the median cottonwood sapling height
66 remained the same or decreased, and many willows and young cottonwoods were intensively
67 browsed (Beschta and Ripple, 2010).

68 Use of the Lamar Valley by wintering elk has declined since wolf reintroduction, due to
69 lower elk numbers and a decrease in the proportion of the elk population wintering on the east
70 side of the range (White et al., 2010; in press). Meanwhile, bison on the northern range increased
71 from 455 in the summer of 1997 (following the removal of 725 the previous winter), to 2070
72 bison in 2007, the highest count on the northern range in the history of the park (Meagher, 1973;
73 White et al., 2011). Since 1984 bison have congregated in the Lamar Valley in summer as well
74 as winter, and some bison have moved from central Yellowstone to the northern range (Taper et
75 al., 2000; Gates et al., 2005; Fuller et al., 2007). Ripple et al. (2010) hypothesized that the bison
76 increase on the northern range may be part of a secondary trophic cascade, where wolves
77 reduced elk density, thereby releasing bison from interspecific competition, resulting in higher
78 bison densities and greater effects from bison on forage plants. Researchers reported seeing bison
79 browsing in the summer season (Beschta, 2003; Beschta and Ripple, 2010), and found willow
80 height to be inversely related to the density of bison fecal piles (Ripple and Beschta, 2006).
81 Significant browsing in summer on the northern range had not previously been reported, nor has
82 browsing by bison (winter or summer) been regarded as an important factor in the ecology of the
83 area by most researchers, who have generally assumed that bison had little effect on browse
84 plants (Singer et al., 1994; Singer and Norland, 1994; Keigley, 1997).

85 Are bison affecting the growth of willow and cottonwood in the Lamar Valley? Summer
86 browsing can distinguish the effects of bison from those of elk, because elk are scarce in the
87 valley in summer. Also, tall willows may be used to compare browsing between heights
88 accessible only to elk, and heights accessible to both bison and elk (Figure 2). We measured the
89 effects of browsing, differentiated by height and season, to answer three questions regarding
90 willow and cottonwood in the Lamar Valley: 1) are these plants suppressed by browsing, 2) how
91 much browsing occurs in summer, and 3) what proportion of browsing can be attributed to
92 bison?

93

94 1.1 Study Area

95 The northern ungulate winter range in the Greater Yellowstone Ecosystem is comprised
96 of open valleys with steppe and sagebrush-steppe vegetation, bordered by slopes with coniferous
97 forest interspersed with aspen groves (Singer and Norland, 1994; Barmore, 2003; Gates et al.,
98 2005). Willow bushes are present in riparian areas and wet meadows throughout the northern
99 range, but cottonwood trees are limited to the larger river valleys (National Research Council,
100 2002; Beschta, 2005; Beyer, 2006; Beschta and Ripple, 2010). Elk and bison share the winter
101 range with smaller numbers of moose (*Alces alces*), mule deer (*Odocoileus hemionus*),
102 pronghorn (*Antilocapra americana*), white-tailed deer (*Odocoileus virginianus*) and bighorn
103 sheep (*Ovis canadensis*) (Singer and Norland, 1994; Barmore, 2003).

104 Study sites were located in the Lamar Valley, a floodplain of the Lamar River about 1 to
105 2 km wide, extending about 9 km from the area of the confluence with Soda Butte Creek on the
106 east to a small canyon called Lamar Canyon on the west (Figure 1). All study sites were west of
107 the Soda Butte Creek confluence. Willow study sites were in wet meadows on the river
108 floodplain, with an additional site along Oxbow Creek about 20 km west and north of Lamar
109 Valley, near where the creek crosses Grand Loop Road (Figures 1, 3). Cottonwood study sites
110 were within the active channel of the Lamar River, where thousands of cottonwood seedlings
111 grow on gravel bars in the wide, shallow, meandering river bed, flanked by meadows of grasses
112 and sedges on the sides of the river (Figure 4). All study locations were within the winter
113 ungulate range.

114

115 2. METHODS

116 Field data were collected between August 20 and September 9, 2010. Plant measurements
117 were similar for willow and cottonwood, but sampling methods were different. Browsing
118 intensity was measured by the percentage of browsed leaders (browsing rate) in the current
119 summer and previous year, the mean length of leader growth since last browsing (growth-since-
120 browsing), and for cottonwood saplings, the mean spring height. We also noted damage from
121 horning (bison thrashing bushes with their horns), and the height of browse-killed stems, defined
122 as a dead stem with at least three terminal twigs at least one of which was pruned by browsing
123 (Keigley, 1997). Season of browsing and growth-since-browsing were determined by examining

124 plant growth architecture, following Keigley and Frisina (1998) and Keigley et al. (2002).
125 Season of browsing was determined by counting the terminal bud scars on the stem. Growth-
126 since-browsing (Keigley's Live-Dead index) was calculated as the difference between the spring
127 height of a stem (the base of current annual growth) and the most recent browse height of the
128 stem, as indicated by the browsed stub (spring height – browse height = growth-since-browsing).
129 Growth-since-browsing compares the current (spring 2010) height of the plant to the height at
130 which it was previously clipped by browsing. This indicator of growth suppression is
131 independent of the height or age of the plant. A strong positive number indicates plants that are
132 growing larger and not suppressed, whereas a negative or small positive number indicates plants
133 suppressed by browsing, because the new growth is lower than or similar to the previous browse
134 height. This occurs when a stem starts a new leader below a leader that was killed by browsing.

135 *2.1 Willow Methods*

136 There were few tall willows in the Lamar Valley, so it was possible to locate all willows
137 taller than 2 m on the Lamar Valley floor (between Lamar Canyon and the confluence with Soda
138 Butte Creek, to the toe of the slope around the valley), and collect data on growth and browsing
139 for all that met the sampling criteria. Height was measured as the spring height, at the base of
140 current annual growth. Willow sites, both in Lamar Valley and Oxbow Creek, were in flat, wet
141 meadows watered by groundwater. Willows within 20 m of a road, in the active channel of the
142 river, or in areas inundated by recent spring river floods were not included, because these factors
143 could affect accessibility and browsing, and flood damage could obscure browsing effects. Most
144 tall willows had a large canopy, but some had few live stems, or were severely damaged by
145 horning, and these were not included. For comparison to the Lamar Valley, we also measured
146 willows on Oxbow Creek, where summer bison use appeared to be very slight (confirmed by scat
147 counts, see Results). All willows shorter than 1 m within the tall willow sites were also
148 measured, using the same methods detailed above for tall willows. Willow bushes in a clump
149 were sampled as a unit if their canopies merged. In the sampled locations willows did not form
150 continuous thickets.

151 To help distinguish the influences of bison and elk, each tall willow bush or clump was
152 divided into two browsing height zones, a lower zone below 1 m accessible to all ungulates, and
153 an upper zone from 1.5 m to 2 m easily accessible to elk but not bison. A pilot study showed that
154 almost all bison browsing occurs below 1 m (authors' unpublished data). Stems between 1 m and

155 1.5 m are unlikely to be browsed by bison, but could have a small amount of bison browsing;
156 therefore, measures in this middle height zone would be ambiguous as indicators of browsing by
157 bison, and were not used.

158 In each height zone we measured four leaders, for a total of eight leaders per bush.
159 Sampled leaders were representative of those most accessible to browsing ungulates, and leaders
160 that were inaccessible to browsers due to dead stems or other obstructions were not included. For
161 those few stems that had never been browsed, the height at which the stem grew beyond
162 browsing obstructions was substituted for the most recent browse height. We sampled an
163 additional 12 leaders in each height zone to estimate the browsing rate for the current summer
164 (2010) and for the previous year (summer 2009 to spring 2010), so browsing rates were based on
165 16 leaders in each height zone. For each bush we also measured the height of three of the oldest
166 browse-killed stems (Keigley, 1997). Variables were averaged for each bush or clump for each
167 height zone, with 95% confidence intervals (*t* distribution), and compared between the two
168 height zones. Because this comparison was between upper and lower heights on the same plants,
169 topographic site variables were ruled out as confounding factors. Browse rates and growth-since-
170 browsing were also compared between willow sites at Lamar and those at Oxbow Creek; these
171 sites were all within the winter range of elk and bison, and were similar in slope, elevation, and
172 water availability.

173 *2.2 Cottonwood Methods*

174 Young cottonwood seedlings and saplings occurred in dense “stands” in discrete sites on
175 alluvial bars along the Lamar River. These sites were relatively homogeneous in age and density,
176 with hundreds of seedlings distributed in a long band on a gravel bar (Figure 4). Many plants
177 were short (<1 m) and hedged, with a bush growth form. We sampled all stands that were longer
178 than 50 m and with most plants older than 3 years, based on the growth visible above the ground.
179 Each stand was a separate study site and sampling unit. For each site, data collected included
180 length, width, distance from river bank, and height above water.

181 A line transect was placed through the centroid along the long axis of each site, and every
182 5 m the plant nearest to the line was measured. For the shortest site, 75 m in length, the sampling
183 interval was shortened to 2.5 m. If the nearest plant was covered with debris above the base of
184 current annual growth, or had less than 3 years of growth visible, or appeared diseased or dying,
185 the next closest plant was chosen. In addition, the tallest cottonwood bush in each 50 m segment

186 (25 m in the smallest site) was measured, as an indication of the leading edge of growth. For
187 each plant we measured the leader with the tallest spring-time height. The field data were used to
188 calculate browsing rate, mean height, mean height of browse-killed stems, and mean growth-
189 since-browsing for each of the seven sites. These quantities were compared using 95%
190 confidence intervals (*t* distribution) to ascertain significant differences among sites, and between
191 height and browse-killed stem height within the same site. The relationship between mean plant
192 height and height above water was analyzed using simple linear regression, to assess the possible
193 influence of water availability. In Site 6, where cottonwood saplings were taller with a single-
194 stem growth form (Figure 4B), the browse status and height for previous years were also
195 recorded.

196 *2.3 Indications of Ungulate Use*

197 The amount of use the study sites received by bison, elk or other ungulates was evaluated
198 based on counts of fecal piles, along with other evidence such as the presence of tracks, wallows
199 and hair, and sightings of the animals. Fecal piles were counted in plots (belt transects) 2 m wide,
200 extending for the length of the wet meadow or cottonwood site. For willows, plots were spaced
201 10 m apart to the edge of the wet meadow containing the willows. For cottonwood, there were
202 two plots in the site and two on the adjacent bank, separated by 4 m. Fecal piles were categorized
203 as from the current summer or a previous season, as determined by color, state of decomposition,
204 and relationship to growing vegetation.

205

206 3. RESULTS

207 *3.1 Willow*

208 Of 53 tall willow clumps found in wet meadows on the floor of the Lamar Valley, 18
209 were rejected because of extensive horning damage (almost all had some horning damage), and
210 three were rejected because their few leaders were protected from browsing by dead branches.
211 Some tall willows growing along the river bank near the east end of the valley were excluded by
212 the decision to limit sampling to wet meadows. The sampled willows included 20 tall willow
213 bushes or clumps in the largest wet meadow and 12 from five additional locations, for a total of
214 32 in the Lamar Valley (Figure 3A). The largest clump was 8.6 m by 4.3 m, the smallest 1.7 m
215 by 0.5 m (the widest extent of live branches). The Oxbow Creek site contained 14 tall willow
216 clumps that met the sampling criteria (Figure 3B). Unlike Lamar, none were rejected due to

217 horning damage and all had full canopies with many leaders. Height ranged from 2.2 to 5.2 m in
218 Lamar (mean 3.5, standard error 0.1), and from 2.3 to 4.2 m in Oxbow (mean 3.1, standard error
219 0.1). All tall willows and most short willows sampled were Geyer willow (*Salix geyeriana*);
220 some short willows were Booth (*S. boothii*) or Bebb (*S. bebbiana*) willow species.

221 Measures of browsing intensity at low height in Lamar Valley were significantly different
222 (*t* test, 95% confidence) from the upper height in Lamar, and also different from either height in
223 Oxbow. These differences were very pronounced (Figure 5). Differences in browsing intensity
224 between upper and lower heights in Oxbow were small, but still statistically significant for
225 growth and previous year browse rate.

226 In the Lamar Valley willow sites, short willows far outnumbered tall willows (196 short/
227 32 tall), but in Oxbow there was a much smaller proportion of short willows (26 short/ 14 tall).
228 For short willows, the summer browsing rate was 88% in Lamar and 0% in Oxbow. Previous
229 year browsing was 100% and 72%, respectively; growth-since-browsing was -3 cm and 7 cm.

230 3.2 Cottonwood

231 There were seven cottonwood sites that met the sampling criteria (at least 50 m long with
232 most saplings older than 3 years) along the Lamar River from near the confluence with Soda
233 Butte Creek to the beginning of Lamar Canyon (Table 1). These seedling patches ranged in
234 length from 75 m to 250 m, and in width from 11 m to 52 m. All were in the active channel of
235 the river, and drift accumulations indicated that four of the seven sites were flooded in the spring
236 of 2010. Most plants were in the form of small bushes (Figure 4A), an indication of intensive
237 browsing (Keigley 1997), and the mean summer browsing rate was 54%. Plants were generally
238 shorter than 1 m except in Site 6 (Figure 4B), but even there the mean spring height was shorter
239 than 1 m (Table 1, Figure 6). Growth-since-browsing was strongly and inversely correlated with
240 summer browsing rate, with both variables log transformed in a linear regression ($r^2=0.92$,
241 $p<0.001$, $n=7$); height was also strongly correlated with summer browsing ($r^2=0.69$, $p<0.02$,
242 $n=7$). Most of the top leaders were browsed in the summer of 2010 preventing direct
243 measurement of current annual growth and productivity, but all of the sites were in a similar
244 landscape position in the active river channel, and there was no significant relationship between
245 mean spring height and height of the plants above water (Table 1; linear regression, $r^2=0.03$,
246 $p=0.70$). Mean spring height was not significantly different (*t* test, 95% confidence) from the
247 mean height of browse-killed stems, except in Sites 1 and 6 where the summer browse rate was

248 low, and this difference was much greater in Site 6, with the lowest summer browse rate. Of the
249 selected saplings in cottonwood Site 6, 19% were too damaged by horning to be measured for
250 browsing, so the next closest sapling was used.

251 *3.3 Ungulate Use*

252 Bison fecal piles were abundant in sampling plots in the Lamar Valley (Table 2), along
253 with wallows and many bison tracks, horned bushes, and clumps of bison hair; there were many
254 bison in Lamar, sometimes browsing willows or cottonwoods (Figure 2). In contrast, no elk
255 pellet piles or tracks from the 2010 summer season were found in either the Lamar Valley or the
256 Oxbow Creek study sites. All bison fecal piles counted in sample plots were found in Lamar,
257 none in Oxbow. The difference in total number of scat piles counted for bison compared to other
258 ungulates was very large in Lamar Valley (Table 2). In both Lamar and Oxbow some elk and
259 bison scat piles from previous seasons were present in the area.

260

261 4. DISCUSSION

262 Willows and cottonwoods in the Lamar Valley were browsed at a high rate, and much of
263 this browsing occurred in the summer season, when herds of bison were present and elk were
264 scarce. Most browsing occurred at low height, and browsing rates were much less at heights
265 above the reach of bison (Figure 5). The season, height, and rate of browsing demonstrate that
266 browsing by bison in summer was common, and that bison were responsible for a large
267 proportion of annual browsing of terminal leaders, enough to suggest a significant ecological
268 effect. For both cottonwood and willow, high summer browsing rates were associated with
269 severely restricted growth (Figures 5, 6). For tall willows in the Lamar Valley, mean growth-
270 since-browsing was negative (-3 cm) at heights below 1 m, showing that stems have not grown
271 back to the heights at which they were previously browsed, a characteristic of bushes that are
272 severely hedged by browsing (Keigley and Frisina, 1998; Keigley et al., 2002). On the same
273 bushes, growth-since-browsing was strongly positive (36 cm) at heights from 1.5 to 2 m,
274 demonstrating that these same willows have been increasing in size at heights accessible to elk
275 but not bison. Similarly for willows shorter than 1 m in the Lamar Valley sites, growth-since-
276 browsing was -3 cm, indicating suppressed growth.

277 If elk were primarily responsible for browsing willows in the Lamar Valley, then the
278 browsing rate would likely be similar in the lower part of a bush and the upper part, because elk

279 can reach the entire height range. Also, the browsing rate should be very low in summer, because
280 the study locations are in elk winter range, with few elk in summer. The summer browsing rate
281 in Lamar was nearly zero in the upper height zone, but very high, 84%, at low height below 1 m
282 (Figure 5). The high summer browse rate is strong evidence that bison are eating most of the
283 accessible leaders before the end of the summer. Browsing below 1 m was also very intensive in
284 the previous year (summer 2009 to spring 2010), with 100% of sampled leaders browsed, as
285 compared to 28% above 1.5 m. Short willows had similar browsing rates, 88% in the summer.
286 The severely hedged condition of willows in the low height range (Figure 3A), and the negative
287 growth-since-browsing (Figure 5), indicate that the high browsing rate measured in the summer
288 of 2010 may represent the typical browsing intensity for recent years.

289 For cottonwood, measurements were compared across the seven cottonwood sites (Figure
290 6). Four sites had summer browsing rates greater than 50%, and two were greater than 90%, with
291 an average of 54%. Growth was suppressed in six of the seven sites as indicated by short average
292 height (<1 m), hedged growth form, and low growth-since-browsing. Only Site 6, with summer
293 browsing rate of 13%, had saplings close to 2 m in height (Table 1, Figure 4). This site was
294 farther out in the river channel than the other sites, and was shaded in winter by a tall adjacent
295 slope, factors that may have reduced browsing and allowed cottonwood saplings to grow taller
296 once pressure from elk was reduced.

297 There was no evidence of elk in the Lamar Valley in summer, either from pellet counts,
298 field sightings, or other evidence, and the area is not considered part of elk summer range. Elk
299 pellets from any season were rare; only 1 elk pellet pile was found in 12,620 m² of scat sampling
300 plots (Table 2). Although detectability was poor in many of these plots, the low elk pellet density
301 is consistent with a major reduction in elk use of the eastern portion of the northern range over
302 the last decade, as reported by White et al. (2010; in press). It was probably during this period of
303 declining elk density that the tall willows in the Lamar Valley grew beyond the reach of elk to
304 their present height; tall willows were not reported in the area previously (Kay, 1990; Chadde
305 and Kay, 1996; Ripple and Beschta, 2006; Beyer et al., 2007). This increased height of willows
306 is evidence of a trophic cascade from wolves to plants; if the increase in bison density is a
307 response to reduced elk density, then the bison increase and their resulting effect on plants would
308 represent an additional pathway associated with this trophic cascade (Ripple et al., 2010).

309 No evidence of moose or deer was found in the Lamar Valley study sites. At Oxbow
310 Creek, deer trails and bedding areas were present among willows, yet summer browsing was
311 minimal, 0.5% percent. Given the clear evidence of deer in Oxbow Creek with little summer
312 browsing, and the lack of any evidence of deer in the Lamar Valley where browsing rates were
313 very high, it is reasonable to conclude that deer were not responsible for the summer browsing of
314 willow and cottonwood observed in the Lamar Valley. Pronghorn were present in the Lamar
315 Valley in summer, in much smaller numbers than bison (Table 2). Studies of the diet and habitat
316 selection of pronghorn on the northern range and elsewhere have found little evidence of willow
317 or cottonwood consumption (Singer and Norland, 1994; Barmore, 2003; Jacques et al., 2006;
318 Boccadori et al., 2008). Low numbers and dietary preferences make it unlikely that pronghorn
319 are having a significant effect on growth or browsing rates of browse plants.

320 Where plants are intensively browsed in summer, as in Lamar Valley, tall willows are
321 constricted below the height at which they are accessible, creating a clump with a mushroom
322 shape. The lower stems are continually clipped but the upper stems continue to lengthen (Figures
323 2A, 3A). This shape, called “highlining,” is seen in many of the tall willows in the Lamar Valley,
324 but where bison are less numerous, as at Oxbow Creek (Figure 3B), willows become full with
325 new growth in summer and have a roughly hemispherical shape. In the Lamar Valley, the low
326 height of this growth suppression suggests that bison, not elk, are now the primary browsers.

327 All tall willows had browse-killed stems with browse brooms (clusters of browse-killed
328 twigs), an indication of past suppression of growth. In the Lamar Valley the mean height of these
329 stems was 147 cm (SE=3.6), and in Oxbow 108 cm (SE=8.3). The fact that these plants are now
330 growing well beyond this previous growth limit is further evidence that they have experienced a
331 release from elk browsing (Keigley and Frisina, 1998). They were previously suppressed by elk,
332 but now are growing freely at those heights.

333 In six of the seven cottonwood sites, the spring height was very close to the height of
334 browse-killed dead stems (Figure 6). This strongly suggests cottonwood saplings at these sites
335 are stunted by browsing, limited to about the same height as the old leaders killed by browsing.
336 The exception is cottonwood Site 6, where live leaders were much taller than browse-killed
337 stems, indicating that something has changed about the browsing and growth dynamics at this
338 site (Figure 4B). In the previous year (summer 2009 to spring 2010) in Site 6, stems shorter than

339 1 m were browsed at a rate of 45%, while those taller than 1 m, above the reach of bison, were
340 not browsed at all, suggesting that recent browsing has been due to bison and not elk.

341 The results of this study make possible an evaluation of alternative explanations for the
342 fact that willow and cottonwood growth in the Lamar Valley has been generally less than in
343 some adjacent areas of the northern range, such as the upper Lamar River and Soda Butte Creek
344 (Ripple and Beschta, 2003; Beschta and Ripple, 2010). One hypothesis could be that there has
345 been no trophic cascade sufficient to release plants from elk browsing. The pronounced changes
346 in height and cover of willow and cottonwood in areas peripheral to the Lamar Valley in
347 conjunction with the recent decline in elk density make this “no-effect” explanation unlikely. An
348 alternative hypothesis is that wolves have caused a release of vegetation by reducing elk
349 browsing, but bison are having an increased effect on plants, counteracting the reduced effects of
350 elk – a secondary trophic cascade (Ripple et al., 2010). The evidence from this research supports
351 this second explanation, for three reasons: 1) plants grew larger and taller where they were
352 beyond the reach of bison, demonstrating release from the effects of elk; 2) browsing rates were
353 very high in summer, when elk were absent, therefore, elk could not have been responsible for
354 most of the browsing of new leaders, because bison consumed them first; and 3) plant growth
355 was suppressed by browsing where the summer browse rate was high, showing that browsing by
356 bison has been affecting plant growth. This growth suppression, and the fact that browsing rates
357 for the previous year were high, are evidence that the summer browsing rates observed in 2010
358 are indicative of a multi-year pattern. The comparison between the lower portion and the upper
359 portion of the same willows shows that differences in site moisture or productivity were not
360 significant factors, as does the similarity in landscape position of cottonwood sites.

361 The bison of Yellowstone today differ from their pre-settlement ancestors in two
362 important ways. First, bison are prevented from moving freely or expanding their range outside
363 the park (White et al., 2011). Second, bison in Yellowstone experience very low predation
364 pressure, compared to what was likely in the past with hunting pressure from humans, and larger
365 numbers of wolves focused on bison (Young and Goldman, 1944; Carbyn, 2003; Kay, 2007).
366 Even if predation was compensatory in ancient times and bison numbers were high, it is likely
367 that predation pressure would have caused bison herds to move, perhaps long distances, as
368 occurs with Canadian bison and wolves (Carbyn, 1997). These differences – freedom to move
369 and greater predation pressure from humans and wolves – make it unlikely that bison would have

370 concentrated in the Lamar Valley in the past as they do today, even if they were present in the
371 region in similar numbers.

372 The consequences of preventing bison movement may extend beyond the bison
373 population to the ecology of the range, in summer as well as winter. The potential effects of
374 bison and other large ungulates include suppression of woody plants and changes in plant
375 communities (Meagher, 1973; Coppedge and Shaw, 1997; Baker, 2003; Gates et al., 2010;
376 Martin et al., 2011). Bison have the potential to limit recovery of willow and cottonwood in the
377 Lamar Valley, and possibly elsewhere in the Yellowstone area. Lack of willow and cottonwood
378 could slow or prevent colonization by beaver and other species, with cascading effects on plant
379 communities, stream morphology, and biodiversity (Kay, 1994; Smith and Tyers, 2008; Baril,
380 2009; Beschta and Ripple, 2010, 2011). Bison, cottonwood, willow and beaver evolved together,
381 but the effects of bison may be more pronounced in Yellowstone today, where bison occur at
382 higher densities and with less movement than was likely when they and the people and other
383 predators that hunted them roamed freely across the landscape.

384

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391

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530 **Tables**

531 **Table 1.** Mean data values for seven cottonwood sites in the Lamar Valley.

Site	Browse Rate (%) Summer	Spring Height (cm)	Browse-killed Height (cm)	Growth Since Browse (cm)	Three Tallest Height (cm)	Height Above Water (cm)	Count
1	23	46	39	6.2	60	138	44
2	68	41	37	-1.4	78	109	44
3	92	26	23	-0.7	40	77	36
4	56	27	27	-0.9	37	73	55
5	27	56	51	6.2	121	75	30
6	13	94	49	35.4	199	74	32
7	100	40	43	-1.7	68	129	30
Grand Mean	54	47	38	6	86	96	39

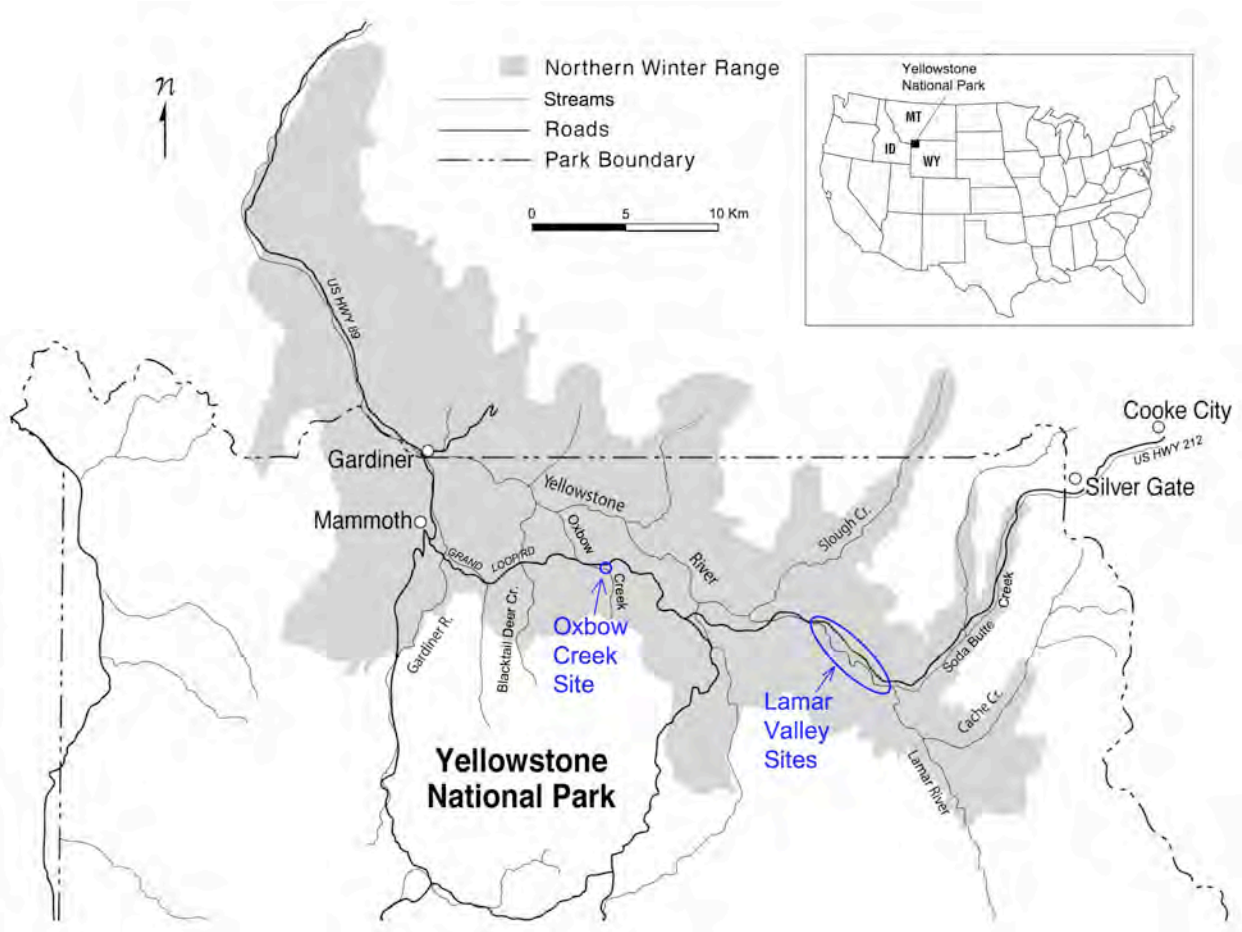
532

533 **Table 2.** Lamar Valley ungulate scat counts, for plots covering 12,620 m². In Oxbow, no scat
534 piles were found in sample plots, but elk and bison scat from previous seasons were present near
535 the site.

Species	Total Fecal			Density (100 m ²)
	Piles	Summer 2010	Older	
Bison	1302	1079	223	10.3
Elk	1	0	1	0.0
Pronghorn	23	23	0	0.2

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540 **Figure 1.** Map of northern Yellowstone National Park, showing the location of study sites at
541 Lamar Valley and Oxbow Creek.
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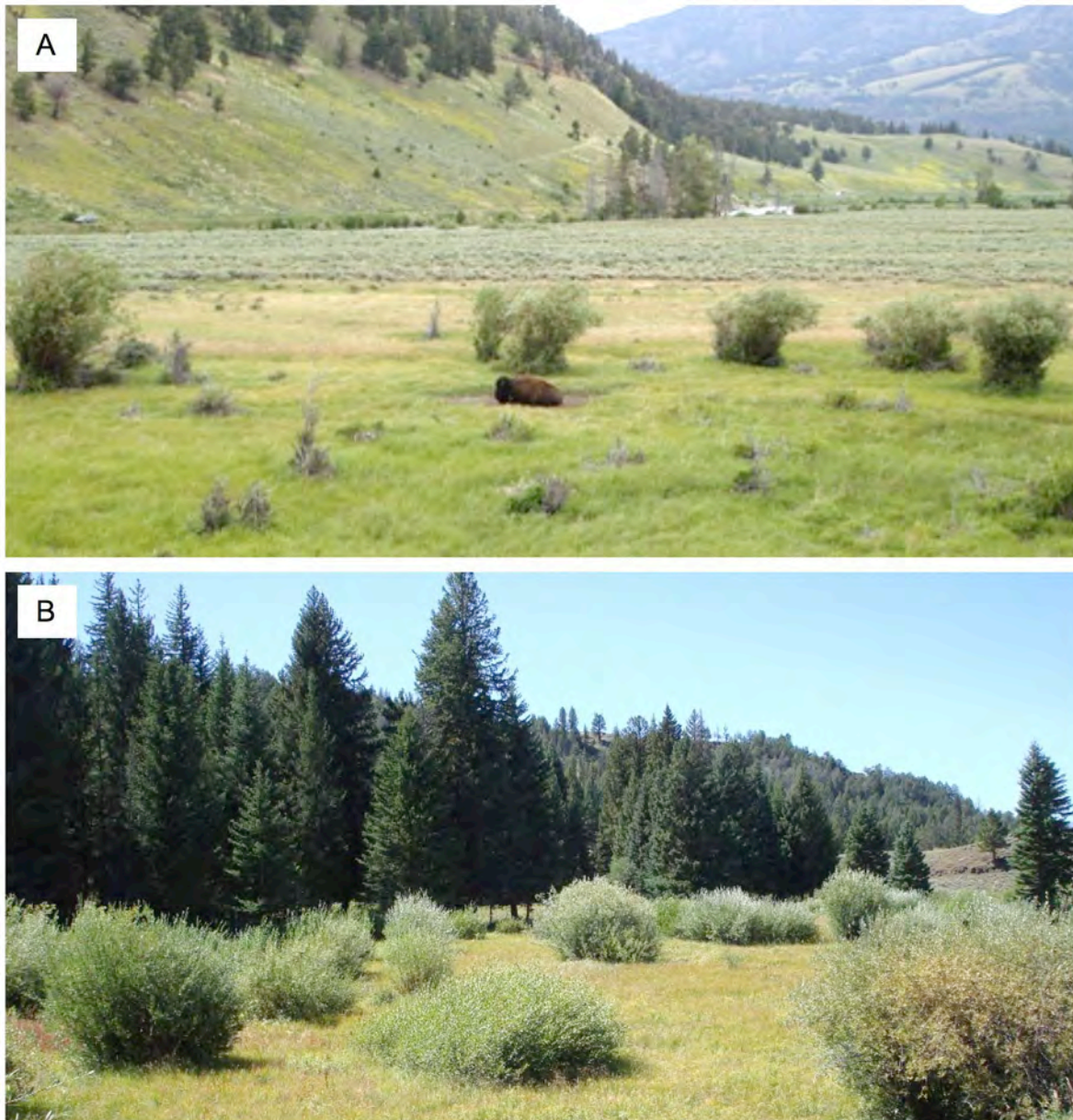
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544 **Figure 2.** Summertime browsing by bison of (A) a willow clump in the Lamar Valley;

545 (B) young cottonwood plants on the bank of the Lamar River. Photos from August

546 2010.

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549 **Figure 3.** Tall willow sites on the Yellowstone northern ungulate range.

550 (A) In the Lamar Valley a bison rests in a willow among tall willows. Willow growth is
 551 constricted by browsing below about 1 m, but expanding above that height, resulting in
 552 a mushroom shape. Most willows are short, with many dead branches (as in
 553 foreground). Tallest willows in the photo are approximately 5 m in height.

554 (B) On Oxbow Creek, willow growth at low height is not suppressed and willows have
 555 a full, hemispherical shape. Most are tall, few are short. Tallest willows in the photo are
 556 approximately 4 m in height. Photos from August 2010.

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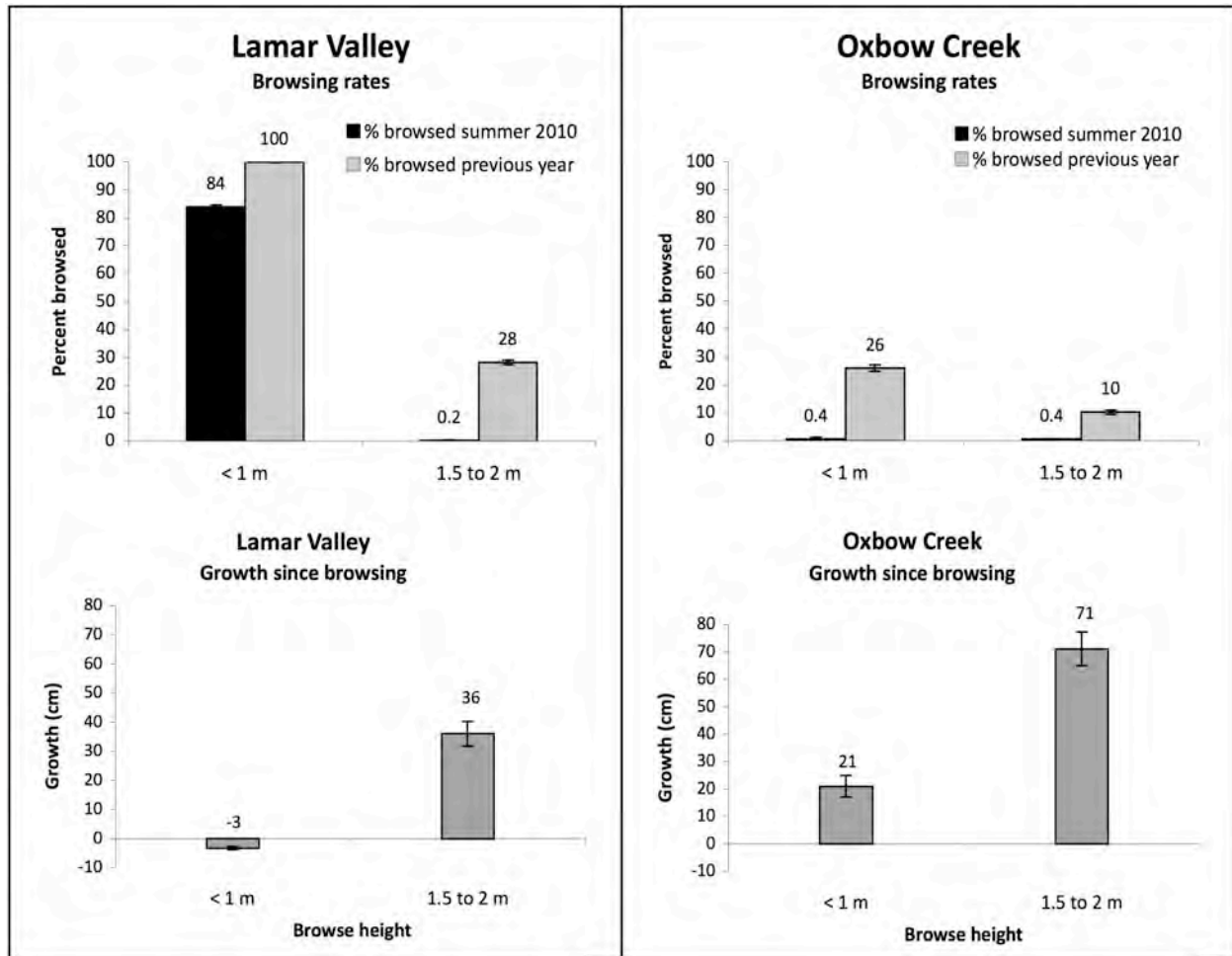
558

559 **Figure 4.** Cottonwood saplings at two sites in the Lamar Valley.

560 (A) Cottonwood Site 4, typical of sites in the Lamar Valley, with hundreds of saplings
 561 hedged and stunted by browsing.

562 (B) Cottonwood Site 6, the exception in Lamar Valley with lower browse rates and
 563 taller saplings. Inconvenient location away from foraging areas, shading by the adjacent
 564 slope in winter, and flooding in spring are possible factors reducing browsing in this
 565 site. All browsing in 2009-2010 was at heights below 1 m. Photos from September
 566 2010.

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570 **Figure 5.** Willow browsing rate and growth-since-browsing (spring height – browse

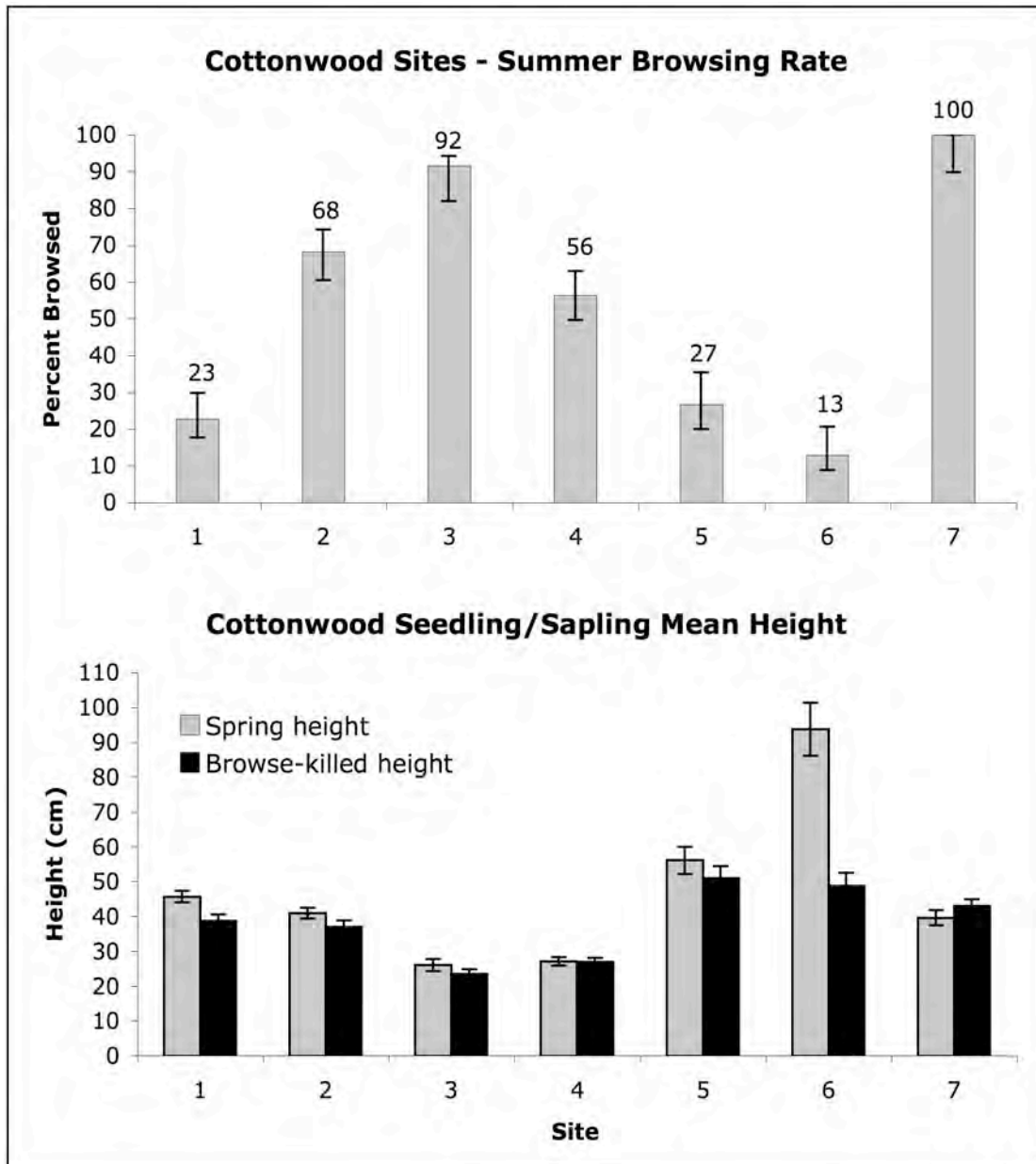
571 height = growth-since-browsing).

572 Summarized data show browsing rate is very high in Lamar below 1 m, but low

573 otherwise, and this difference is reflected in growth (bars show standard error). Summer

574 browsing is near 0% except in Lamar Valley at low height.

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Figure 6. Seven cottonwood sites in the Lamar Valley in summer 2010 (bars show standard error). A) Four of seven sites had summer browsing rates >50%, with two greater than 90%. B) Mean height was strongly suppressed except in Site 6, where summer browse rate was very low. Mean height was similar to the height of browse-killed stems except in Site 6, so plants have grown little beyond the height at which they were previously hedged by browsing except where summer browse rate was low.